

Spatial Analysis of Groundwater Sources in the Duport Road (Shara and Cow Field) and Soul Clinic Diamond Creek communities, Paynesville City, Republic of Liberia

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ABSTRACT

Groundwater is of significant water source use for drinking in sub-Sharan nations. About 90% of the family in the study area rely extensively on groundwater for drinking. The study area is selected communities in the Paynesville City, Greater Monrovia. The selected communities are the Duport Road Shara and Cow Field, and the Soul Clinic Diamond Creek communities. Ordinary Kriging and the Global Moran's Index methods were used to map the spatial distribution and dependence of water parameters present in the water infrastructure. The water infrastructure for this research defined as hand pump/borehole and opened and covered shallow wells. Results from the study showed 93% of the groundwater water infrastructures contaminated with total coliform bacteria. The distribution of the physical and chemical parameters in the groundwater reduced towards the north of the study area, the Soul Clinic Diamond Creek community. From the ordinary Kriging method, the nitrated distributed concentration and temperate has moderate to high spatial dependence, while peroxide, total dissolved solids, total alkalinity, and pH have weak spatial dependence. The Global Moran's Index results show that nitrate distribution and the temperature was statistically significant. Finally, the nitrate concentration is a public health and environmental issues. Finally, the water infrastructure in the community is a serious public health problem and need immediate attention. Most of the water infrastructure not suitable for drinking.

Keywords : Duport Road, Greater Monrovia, Paynesville, Ordinary Kriging, Global Moran's Index, groundwater, total coliform, microbial, chemicals, physical

I. INTRODUCTION

Groundwater and aquifers are complex and changeable ecosystems of grave importance for geochemical cycles [1]. Shallow wells are the primary source of collecting groundwater in many rural and peri-urban communities. Shallow wells are vulnerable to fecal contamination, which is often due to leaching pit latrines [2], sewage and surface runoff. Areas where the prevalence of open defecation is high, groundwater may be exposed to fecal contamination. The decrease in groundwater quality attributed to an increased contribution of non-point pollution sources (NPS), with strongly dependent on land use[3].

Human activities associated with water consumption or unsustainable groundwater depletion is a widespread practice across the globe [4], especially in developing nations. Groundwater quality compares to surface water is of good quality and better protected against contamination[5]. Groundwater in the study areas experiences seasonal variation in the volume of the water source. During the rainy season from mid-May to mid-October, groundwater level increased and in swamp area exceed the water table level. Improvement of water sources associated with the reduction global burden disease. In developing nations, about 80% of sickness is related to drinking polluted water[6].



Figure 1: Left: Location of the Study Area in Montserrado County. Right: Map of the Republic of Liberia with the Montserrado County

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Groundwater is a fundamental source of drinking water in Sub-Saharan Africa [7]. However, the ecoenvironmental problem caused by the exploitation and utilization such as rising groundwater level associated with groundwater quality problem[8]. From Environmental Health perspective, human activities and developmental purposes increased population exposed to groundwater contaminant. Surface water and or groundwater pollution is related to increased population, economic growth, and industrialization [9].

Solid waste management and sewage disposal or open defecation are significant to the quality of groundwater in the study areas. In the study, the prevalence rate of open defecation is 28%, and water sources not adequately managed. Therefore, the objective of the survey is to spatially determine the path of the contaminant (chemical, physical, and microbial) in the water infrastructure.

II. METHODS AND MATERIAL

Groundwater contamination is a global challenge due to rapid industrialization and population growth [10]. The groundwater sample was collected between the month of May to August 2016 from the Soul Clinic and Duport Road communities in Paynesville City, Montserrado County, Republic of Liberia. The water sample was collected from protected and unprotected water sources from May 2016 to August 2016 to obtain the water parameter. The water sample parameter measured from the chemical, physical and biological contaminants in the water sample. The water sample collected from 43water infrastructure. Parameters tested included; chemical (include nitrate, nitrite, peroxide, metals), physical (TDS, Total Hardness, Total Alkalinity) and the biological and related parameters include total coliform bacteria, pH, temperature.

Study Area. The study area located in Greater Monrovia, Republic of Liberia. The Republic of Liberia situated on the West Coast of Africa with a land cover of 43000 square miles. The climate is tropical. The study areas focused on selected communities in Paynesville City. In the selected communities-Duport Road and Soul Clinic, approximately 90% of families rely on groundwater for drinking and domestic work. Groundwater in the study collected from a constructed water infrastructure. The water infrastructures in the study areas are hand pump and opened and covered shallow wells.

Parameter. The chemical and physical parameters were analyzed directly on the field using HACH DR 900 handheld equipment. For the microbial analysis (total coliform bacteria) the LaMotte total coliform test kit was used. The theory of the test kit based on presence and absence of total coliform bacteria. The GPS coordinate was logged using the GARMIN ETREX 10 GPS.

Distribution of Wells/Hand Pump in the Study Area



Figure 2: Distribution of Wells/Hand Pump in the Study Area

III. RESULTS AND DISCUSSION

The analysis process and prediction of the water parameters in the study area done using geostatistical (Ordinary Kriging) and spatial analysis (Global Moran' Index) tools. For the geostatistical analysis, normalization was fundamental for with high skewness or kurtosis using the log transformation (Table 1). The geostatistical and spatial analysis was done using ArcGIS version 10.4.

Descriptive

The descriptive results generated from the geostatistical-Ordinary Kriging tools. Table 1, depicts descriptive statistic summary of the groundwater parameters obtained from the water sample. Table 1 gives the difference between the skewness and kurtosis before and after transformation for each parameter.

Table 1: Descriptive Analysis of the groundwater(Wells/Hand Pump) Hydrological Parameters

No	Hydrological Parameters	# of Data	Min	Max	Mean	St. Dev	Skewness	Kurtosis
1	TDS	43	14	590	89.3	99.2	3.42	16.9
	TDS ^a	43	2.64	6.38	4.13	0.82	0.46	2.84
2	PEROXIDE	43	0.3	25	7.68	7.01	1.32	4.13
3	pН	43	3.4	6.4	5.3	0.81	-0.62	2.43
	pHª	43	1.22	1.86	1.66	0.16	-0.88	2.90
4	TH	43	25	425	172	101	1.33	4.09
	TH ^a	43	3.22	6.05	4.99	0.58	-0.32	4.14
5	TA	43	0	720	73.2	117	4.14	22.8
6	NITRATE	43	0.23	80	15.9	16.2	2.08	8.11
7	TEMP	43	26	32	29.1	1.67	0.21	2.36
	TEMP ^a	43	3.26	3.47	3.37	0.057	0.097	2.38

^aTransformation using logarithm

Groundwater Classification in the Study area

The groundwater is classified based on the analysis of samples collected from the water infrastructure. The water sample was analyzed in the field using the HACH DR 900 handheld equipment for the chemical and physical parameters. From Table 1, the concentration of the water parameters from the 43- water infrastructure varies. Figures 3 to 6 provide information on the distribution of the water parameters from geostatistical analysis using the Ordinary Kriging tool. Figure 3 gives the distribution of the water infrastructure in the study areas.

Nitrate & Peroxide. Figure 4 shows the distribution of nitrate and peroxide. The concentration of nitrate (Table 1) varies from 0.23 to 80ppm with the average concentration of 15.9 ± 16.2 ppm. From the WHO 2017 drinking water guidelines, nitrate accepted range is 50ppm. Therefore some of the water infrastructures were above the allowed range of nitrate in drinking water. 6.98% approximately 7% (coordinates: 6.268167, -10.660633 (Shara); 6.273983, -10.671867 and 6.274133, -10.671667 (Cow Field)) of the water infrastructure constructed between 2010 – 2015 was above the WHO

drinking water guidelines for nitrate. The type of water infrastructure was two opened or unprotected well and a covered or protected well.

The nitrate and peroxide concentrations reduced to the north of the study area and distributed randomly central of the study area. However, the concentration of the peroxide range (Table 1) between 0.3 to 25ppm and the average peroxide concentration is 7.68 ± 7.01 ppm.

Nitrate a univariate anion interfere with the intake of iodine by thyroid resulting to the reduced production of thyroid hormones. In previous epidemiological studies, investigators a correlates nitrate contamination in water supplies[11] and thyroid dysfunction and thyroid disease. In a cross-sectional study of school children l in areas of Slovakia with increased levels of nitrate concentration exposure via drinking water, experience increased thyroid disease [12]. The presence of nitrate in groundwater where the used of fertilizer may be rare could be associated with decaying organic matter and sewage[13] leakage.

Nitrate itself is a compound of low toxicity but reduced to nitrite, the toxic form relating human health, [14] which can cause infantile methemoglobinemia. The reduction reaction is biochemical in the human system that released an oxygen atom from the nitrate (NO₃) compound to produce nitrite (NO₂).

Total Dissolved Solids & Total Hardness. Figure 4 shows the distribution of the total dissolved solid and total hardness in the water infrastructure in the study area. The total dissolved solids (TDS) vary between 14 - 590ppm with the average concentration of 89.3 ± 99.2 ppm. The total dissolved solids reduced towards the north of the study area, the Soul Clinic Diamond Creek community. However, 6 (14%) of the 43-water infrastructure constructed between 2006 – 2014 in the Shara and Cow Field communities was above the WHO drinking water guidelines. The six WI include one hand pump in Shara and four – open well, one covered well in the Cow Field community. The accepted total dissolved solids in drinking water are 150ppm.

The distribution of total hardness in the study area was randomly distributed (figure 4). The total hardness concentration varies from 25 to 425ppm with average means of 172 ± 101 ppm. 4 (9%) of the water infrastructure above the WHO drinking water guidelines. The drinking water guideline for total hardness from the WHO 2017 guidelines is 270 ppm. The four water infrastructures are in Shara (one-hand pump and two opened well) and Cow Field community (one-opened well) constructed between 2003 to 2015.

pH and Total Alkalinity. Figure 5 depicts the distribution of the acidity and alkalinity of the water sample from the 43-water infrastructure in the study area. The distribution shows that the water was more acidic associated with the dissolution of metal in the soil. The pH range between 3.4 to 6.4 and from the accepted WHO 2017 drinking water guidelines, the pH range is 6.8 to 8.5. None of the water source meets the WHO guideline for pH level in drinking water. The average pH value is 5.3 ± 0.81 . For the total alkalinity, about 95% of the water infrastructure was between 0 - 180 ppm with the highest value of 720ppm from a hand pump in Shara community. The average of the total alkalinity was 73.2 \pm 117ppm. The pH is a critical parameter for predicting the amount of dissolved CO₂, the precipitation of carbonates in the pore spaces, and the release of trace metals from the reservoir rocks [15]. Contaminated groundwater typically has chloride and alkalinity levels ranging from micro- to millimole, and pH from neutral to extremely high concentrations impacted by the base activation of persulfate. Alkalinity and pH mainly affect the surface complexation of containing aquifer soluble minerals.

Higher alkalinity favors the formation of unreactive surface carbonate complexes, while higher pH favors the formation of reactive surface hydroxy complexes and accelerates remediation efforts [10]. In an acidic media, both reduction and oxidation of Fe microbially catalyzed, and available evidence suggests that microbial Fe redox cycling takes place across a diverse wide range[16] of modern natural environments.

Total Coliform Bacteria & Temperature. Figure 6 shows the distribution of total coliform bacteria and related temperature value from the water source. The total coliform bacteria were positive in 93% of the 43 water infrastructures water sample tested. The total coliform analysis for the study was based on the qualitative analysis using the either presence or absence

of total coliform bacteria in the water sample. The total coliform bacteria was determined using LaMotte total coliform kit. The average temperature was $29.1 \pm 1.67^{\circ}$ C and range from 26 to 32° C. A study conducted in Nigeria showed that 85% of wells tested was acidic that is below WHO standard for pH[9]. The study also had similar results with all the 43-water sources pH range was between 3.3 to 5.5.

Drinking water contamination risk reflects the increased vulnerability of groundwater sources to surface contamination, particularly in flood area. A study conducted in Bangladesh revealed fecal contamination due to flood. Tube well water samples in the flooded areas were contaminated with total coliforms (41%, n = 85), thermotolerant coliforms (29%, n = 60) and E. coli (13%, n = 27). Only four samples had E. coli contamination >100 CFU 100 ml) [17]. The study area does experience flood during the rainy season affecting groundwater infrastructure. Relatively high groundwater table within flood areas exposed to fecal organisms.

Maintaining a high-quality drinking water is one of the fundamental goals of health or public health authorities in many countries. Thus, the establishment of the legislature of the drinking water quality guideline, in developed and developing nations aid in the production of quality water for drinking and domestic work [4].

Geostatistical Analysis

Figure 7 depict scatter plots or semi-variogram around the model indicated by blue line or average. Table 2 illustrates the suitable semivariogram models for the hydrogeological parameters from the 43-wells water sample. The range from table 2, is the distance where the design flattens and varies among the parameters. The sill and the nugget provide information relating to attainment of range and interception of the y-axis respectively.

The ratio of the nugget to the sill correlates the spatial dependences of groundwater quality. The proportion of the nugget per skill has three classifications: less than 25% high spatial dependence, in the range of 25% to 75% moderate spatial dependencies and greater than 75% weak spatial relationships[18]. From Table 2 nitrate concentration, total hardness, and the temperature showed a moderate to high spatial dependence.



Peroxide(ppm) Distribution



Figure 3: Distribution of Nitrate (Left) and Peroxide (Right) concentration in the study area



Figure 4: Distribution of TDS (Total Dissolved Solids) (Left) and TH (Total Hardness) (Right) concentration in the study area.



Figure 5: Distribution of pH (Left) and TA (Total Alkalinity) (Right) concentration in the study area



Figure 6: Distribution of Total Coliform Bacteria (Left) and Temperature (Right) concentration in the study area



Figure 7 : Fitting Semivariogram models for the Groundwater Hydrogeological parameters, i.e., a=peroxide, b=nitrate, c=Total Dissolved Solids, d=Total Hardness, e=pH, f=Total Alkalinity, g=Temperature

Hydrogeological	# of	Transformation	Number	Lag	Nugget	Sill	Partial	Range(M)	Nugget/Sill	Nugget/Sill
Parameters	Data		of Lag	Size			Sill			
Peroxide	43	Normal	12	0.00038	1.43	1.481	0.051	0.0023	0.966	Weak
TDS	43	Logarithm	12	0.00034	7,521	9,641	2,093	0.0027	0.780	Weak
TH	43	Logarithm	12	0.00025	0.164	0.323	0.159	0.0021	0.508	Moderate
ТА	43	Logarithm	12	0.00022	0.802	1.00	0.198	0.0016	0.802	Weak
Temp	43	Logarithm	12	0.00057	0.001	0.003	0.0021	0.0039	0.333	Moderate
Nitrate	43	Normal	12	0.00022	0.741	1.771	1.03	0.0015	0.418	Moderate
pH	43	Logarithm	12	0.00071	0.022	0.026	0.004	0.0054	0.846	Weak

Table 2 : Semi variogram Characteristics for Map generation

Table 3: Global Moran's Index Analysis for the hydrogeological parameters

Moran's Index	P – Value	Z-Score
0.240	0.000082**	3.94
0.066	0.203	1.27
-0.00027	0.731	0.343
0.032	0.560	0.584
0.059	0.25	1.15
0.376	0.000001**	5.55
0.068	0.11	1.60
	Moran's Index 0.240 0.066 -0.00027 0.032 0.059 0.376 0.068	Moran's Index P - Value 0.240 0.000082** 0.066 0.203 -0.00027 0.731 0.032 0.560 0.059 0.25 0.376 0.000001** 0.068 0.11

** = Statistically Significant

Table 4: Water Quality data from the three selected communities – Soul Clinic Diamond Creek and the Duport Shara and Cow Field communities

СО		LON		IN	НО	TDS	TEMP	Р	TH(pp	TA (pp	PEROX IDE	NO3	Total COLIFOR
DE	LAT	G	СОМ	F	Μ	(ppm)	(0C)	Н	m)	m)	(ppm)	(ppm)	М
W1	6.269 883	- 10.66 44	SHARA	H P	20	590	27	3. 8	120	0	5	15.5	NEGATI VE
W2	6.269 75	- 10.66 36	SHARA	H P	10	130	28	4. 5	120	40	2	19.9	NEGATI VE
W3	6.272 35	- 10.66 27	SHARA	O W	8	31	29	3. 9	125	0	10	20	POSITIV E
W4	6.272 283	- 10.66 17	SHARA	O W	10	42	29	3. 4	250	40	5	20.1	POSITIV E
W5	6.272 017	- 10.66 17	SHARA	H P	15	109	29	6. 1	425	240	10	2.3	POSITIV E
W6	6.271 367	- 10.65 93	SHARA	H P	15	141	29	6. 3	150	720	5	0.98	POSITIV E
W7	6.270 983	- 10.65 97	SHARA	O W	5	68	29	6	425	120	0.6	1.98	POSITIV E
W8	6.270 067	- 10.66 11	SHARA	C W	4	42	28	5. 3	120	0	25	22.9	POSITIV E
W9	6.269 65	- 10.66 13	SHARA	O W	6	52	28	3. 8	120	0	0.5	17	POSITIV E

W10	6.269 4	- 10.66 1	SHARA	C W	10	38	29	5	120	0	5	12.6	POSITIV E
W11	6.268 9	- 10.66 08	SHARA	O W	15	47	29	5. 8	120	40	5	17.8	POSITIV E
W12	6.268 35	- 10.66 14	SHARA	H P	14	18	28	4. 1	50	0	5	20.7	NEGATI VE
W13	6.268 083	- 10.66 14	SHARA	O W	20	92	29	6. 4	250	120	0.5	40	POSITIV E
W14	6.266 15	- 10.65 69	SHARA	H P	5	130	32	5. 8	120	40	10	5.9	POSITIV E
W15	6.265 967	10.65	SHARA	H P	13	.58	29	5.	180	40	5	20.1	POSITIV E
W16	6.265	10.65	SHARA	H P	14	120	31	5.	50	40	10	4.12	POSITIV E
W17	6.265 033	10.65	SHARA	H P	12	42	32	5.	120	40	5	4	POSITIV E
W18	6.264 083	10.65	SHARA	O W	3	150	30	6. 1	250	120	10	0.23	POSITIV E
W19	6.268 217	- 10.65 78	SHARA	H P	17	120	32	5. 7	120	0	0.3	0.67	POSITIV E
W20	6.267 433	- 10.65 98	SHARA	O W	12	150	32	6	120	80	25	1.8	POSITIV E
W21	6.267 033	10.66	SHARA	C W	15	38	29	5. 7	120	40	10	20.5	POSITIV E
W22	6.265 383	- 10.66 15	SHARA	C W	20	36	30	5. 7	120	40	2	3.9	POSITIV E
W23	6.265 333	10.66	SHARA	C W	18	34	28	5.	120	40	5	2	POSITIV E
W24	6.267 25	- 10.66 08	SHARA	O W	10	53	30	6. 2	120	40	10	20.5	POSITIV E
CO DE	LAT	LON G	СОМ	IN F	HO M	TDS (ppm)	TEMP (0C)	P H	TH(pp m)	TA (pp m)	PEROX IDE (ppm)	NO3 (ppm)	Total Coliform
W25	6.266 717	- 10.66 03	SHARA	O W	25	30	27	5. 8	140	80	0.5	20.9	POSITIV E
W26	6.268 167	- 10.66 06	SHARA	O W	18	73	31	6. 4	130	40	5	50.9	POSITIV E
W27	6.269 1	- 10.65 95	SHARA	H P	15	24	32	4. 4	120	0	10	3.4	POSITIV E
W28	6.269 117	- 10.65 81	SHARA	O W	15	40	32	5. 3	120	180	25	5.6	POSITIV E

W29	6.269 767	- 10.65 81	SHARA	O W	13	95		31	6. 4	425	120		10	3.6	POSITIV E
W30	6.269 483	- 10.65 58	SHARA	C W	15	26		30	4. 9	250	80		10	19.4	POSITIV E
W31	6.268 833	10.65 57	SOUL CLINIC	O W	5	79		28	6. 2	250	80		2	19.8	POSITIV E
W32	6.292 35	- 10.66 28	SOUL CLINIC	C W	10	14		28	4. 7	250	40		0.5	4.9	POSITIV E
W33	6.292 833	- 10.66 25	SOUL CLINIC	H P	25	22	26		5. 3	120	40	2		5	POSITIV E
W34	6.292 1	- 10.66 23	SOUL CLINIC	C W	25	26	27		4. 1	120	40	0.4		17	POSITIV E
W35	6.292 267	- 10.66 12	SOUL CLINIC	O W	15	28	27		4. 7	250	80	2		18	POSITIV E
W36	6.292 967	- 10.66 06	SOUL CLINIC	C W	20	28	27		4. 3	120	0	5		10	POSITIV E
W37	6.292 317	- 10.66 27	SOUL CLINIC	O W	10	120	26		5. 4	250	40	10		15	POSITIV E
W38	6.273 583	- 10.67 18	COW FIELD	O W	20	340	29		5. 1	425	120	10		15	POSITIV E
W39	6.273 983	- 10.67 19	COW FIELD	C W	15	153	29		4. 9	120	0	25		80	POSITIV E
W40	6.274 133	- 10.67 17	COW FIELD	O W	20	154	30		5. 7	250	250	10		60	POSITIV E
W41	6.274 617	- 10.67 13	COW FIELD	C W	25	27	29		5. 6	125	80	10		18	POSITIV E
W42	6.275 157	- 10.67 07	COW FIELD	H P	15	140	30		5. 4	75	40			20	POSITIV E
W43	6.252 967	- 10.66 82	COW FIELD	H P	10	88	28		5. 6	25	0	2		20	POSITIV E

Note: LAT = Latitude; LONG = Longitude; COM = Community; INF = Water Infrastructure; HOM = number of Homes per water infrastructure; TDS = Total Dissolved Solids; TEMP = Temperature; TH = Total Hardness; TA = Total Alkalinity; NO3 = Nitrated

Table 5: Number of homes collecting groundwater contaminated with Nitrate and Total coliform bacteria

	Covered	d Shallow /ells	Hand	Pump	Opened Sha Wells	llow	N	litrate		
	Total Coliform Results (ppm)								Number	
	Positive	Negative	Positive	Negative	Positive	Ne gati ve	Min	Max	of Homes	
Diamond Creek/Soul Clinic	3		1		3		80	15	105	
Duport Rd/Cow Field	2		2		2		50.9	0.23	392	
Duport Rd/Shara	10		7	3	10		19.8	4.9	110	

Spatial Autocorrelation (Spatial Analysis)

The spatial autocorrelation or Global Moran's Index is an inferential statistic tool with results interpreted within the context of the null hypothesis. For the Global Moran's Index statistic, the null hypothesis suggests that the water parameters are randomly distributed among water infrastructure in the study areas.

The Moran's Index uses the Getis-Ord General G to verify the null hypothesis. The G-statistic revealed a trend toward a concentration of high values with very high statistical significance in the selected sites indicating spatial autocorrelation [19]. The Getis-Ord statistic gives more natural results and better visual exploration and has the advantage of distinguishing high-value clusters or low value clusters[20].

From Table 3, the Moran's I index for the nitrate concentration and the temperature give a positive value, 0.240 and 0.376 respectively. The Z – score is 3.94 and 5.55 for nitrate and temperature respectively. Therefore, the spatial distribution of nitrate concentration and temperature whether high or low spatially clustered. The standard deviation of both nitrate and temperature is statistically significant at the 0.01 level. Which implies that there is less than 1% likelihood that the observed pattern of the nitrate concentration and temperature could have occurred by chance. The result indicates the observed pattern of nitrate and temperature have a significant impact on the water quality use for drinking in the study areas.

IV. CONCLUSION

From the results, it suggests that the 43-water infrastructure does not meet the drinking water guidelines, thus unfit for drinking. The three water infrastructures (Hand pump, Opened and Covered Wells) were above the accepted nitrate concentration and contaminated with total coliform bacteria located in the Shara and Duport Road - Cow Field communities. The nitrate contaminated water infrastructures served 53 homes with 66% from Duport Road Cow Field community and 33.9% from the Duport Road Shara community. For the distribution of contaminant, the parameters decreased towards the North of the study area; the Soul Clinic Diamond Creek community excepts for total hardness and acidity fluctuating concentration. Regarding chemical and physical parameters, the water sources in the North of the Study area are much better as compared to the Duport Road Shara and Cow Field communities. The three communities water infrastructure are contaminated with total coliform bacteria. Kriging method was used in the mapping prediction of contaminant concentration in the water infrastructure. A log transformation was applied for some hydrogeological parameters to enhance normalization of the concentration. The total dissolved solids, total hardness, pH, and temperature underwent the log transformation while the peroxide and nitrate concentration transformation were not applicable because the skewness and kurtosis were at the acceptable level. The semi variogram analysis varied for each hydrogeological parameters. The hydrogeological parameters concentration of a geological location, the nitrate concentration, and temperature have a moderate

to high spatial dependence. All other parameters have a weak spatial dependence. The global Moran's Index results show that the nitrate concentration and temperature are statistically significant. The results imply that the 99% of the distribution of the nitrate concentration in the study area do not occur by chance may be either associated with environmental activity since agriculture activity that may require extensive use of fertilizer does not exist in the study area. Finally, the study results indicate that the water infrastructure located in the study area needs immediate attention and further research. Results from the Kriging and the Spatial Autocorrelation analysis show that the nitrate contaminant in the water infrastructure is an immediate public health and environmental issues in the study.

V. REFERENCES

- Vigneron, A., et al., Microbial and Isotopic Evidence for Methane Cycling in Hydrocarbon-Containing Groundwater from the Pennsylvania Region. Front Microbiol, 2017. 8: p. 593.
- [2]. Velasquez-Orta, S.B., et al., Microbial fuel cells for inexpensive continuous in-situ monitoring of groundwater quality. Water Res, 2017. 117: p. 9-17.
- [3]. Infascelli, R., R. Pelorosso, and L. Boccia, Spatial assessment of animal manure spreading and groundwater nitrate pollution. Geospatial Health Journal, 2009. 4(1): p. 27-38.
- [4]. Voisin, N., et al., Effects of Spatially distributed sectoral water management on the redistribution of water resources in an integrated water model. AGU Publication, 2017: p. 01-18.
- [5]. Okkonen, J. and B. Klove, Assessment of temporal and spatial variation in Chemical Composition of groundwater in an unconfined esker aquifer in the Cold temperate climate of Northern Finland. Cold Regions Science and Technology, 2011: p. 118-128.
- [6]. Engstrom, E., et al., Pravelence of microbial contaminants in groundwater sources and risk factorin Juba, South Sudan. Science of Total Environment, 2015: p. 181-187.
- [7]. Anormu, G., A. Gibrilla, and D. Adomako, Tracking nitrate sources in groundwater and associated health risk for rural communities in the white Volta River Basin of Ghana using isotopic approach (δ15N, δ18O - NO3 and 3H). Science of Total Environment, 2017.
- [8]. Han, D., et al., A survey of goundwater levels and hydrogeochemistry in irrigated fields in the Karaway Agricultural Development Area, northwest China; Implications for soil and groundwater salinity resulting

from surface water trnsfer for irrigation. Journal of Hydrology, 2011. 405: p. 217-234.

- [9]. Aboyeji, O.S. and S.F. Eigbokham, Evaluation of groundwater contamination by leachates around Olusosun open dumpsite in Lagos metropolis, southwest Nideria. Journal of Environmental Management, 2016. 183: p. 333-341.
- [10]. Li, W., et al., Mechanisms on the Impacts of Alkalinity, pH, and Chloride on Persulfate-Based Groundwater Remediation. Environ Sci Technol, 2017. 51(7): p. 3948-3959.
- [11]. Aschebrook-Kilfoy, B., et al., Modeled nitrate levels in well water supplies and prevalence of abnormal thyroid conditions among the Old Order Amish in Pennsylvania. BioMed Central Journal, 2012: p. 11-16.
- [12]. Aschebrook-Kilfoy, B., et al., Modeled nitrate levels in well water supplies and prevalence of abnormal thyroid conditions among the Old Order Amish in Pennsylvania. Environmental Health Journal, 2012: p. 01-11.
- [13]. Barkouch, Y., E.K.M. Eddine, and A. Pineau, A New Approach to Understanding Well Water Contamination by Heavy Metals at a Mining Extract Region in Marrakech, Morocco. Polish Journal of Environmental Studies, 2016. 25(03): p. 1347 - 1351.
- [14]. Infascelli, R., R. Pelorosso, and L. Boccia, Spatial assessment of animal manure spreading and groundwater nitrate pollution. Geospat Health, 2009. 4(1): p. 27-38.
- [15]. Mito, S., K. Okamura, and H. Kimoto, Colorimetric pH Measurement of Pressurized Groundwater Containing CO2. Anal Sci, 2016. 32(4): p. 437-42.
- [16]. Roden, E.E., et al., The Microbial Ferrous Wheel in a Neutral pH groundwater Seep. Frontiers in Microbiology, 2012. 03(172): p. 1 - 18.
- [17]. Luby, S.P., et al., Tubewell water quality and predictors of contamination in three flood-prone areas in Bangladesh. Journal of Applied Microbiology, 2008. 105 p. 1002–1008.
- [18]. Marko, K., N. Al-Amri, and A.M.M. Elfeki, Geostatistical analysis using GIS for Mapping Groundwater Quality: Case study in the reharge area of Wadi Usfan, Western Saudi Arabia. Saudi Society Journal for Geoscience, 2014(7): p. 5239-5252.
- [19]. Garcia-Palomares, J.C., J. Gutierrez, and C. Minguez, Identification of tourist hot spots based on social networks: A comparative analysis of European metropolises using photo-sharing services and GIS. Applied Geography, 2015. 63: p. 408-417.
- [20]. Ding, L., et al., Spatial-Temporal Hotspot Pattern Analysis of Provincial Environmental Pollution Incidents and Related Regional Sustainable Management in China in the Period 1995-2012. Sustainability, 2015. 7(10): p. 14385-14407.